

Ecological Effects of Lead Mining in Streams of the Missouri Ozarks

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ABSTRACT

The Viburnum Trend extends through the Black River drainage of southeast Missouri and is one of the largest producers of lead ore in the world. Mining interests seek to expand lead exploration in the adjacent Mark Twain National Forest and the Ozark National Science Riverways areas of southeastern Missouri. We studied the structure of fish and invertebrate communities at 12 sites in or near the Black River watershed to determine if existing lead mining activities were affecting aquatic biological communities. Stream surveys were conducted at four reference sites with no known upstream mining activities; eight sites downstream of known mining areas were also sampled. Surface and porewater samples collected in riffle habitats (i.e., coarse substrate) were analyzed for metals, major cations, dissolved organic carbon, sulfate, and ammonia. Physical stream habitat was characterized using microhabitat-level variables (i.e., velocity, depth, substrate composition) as well as landscape-level variables (i.e., watershed area, land use, and channel morphology). Surface and porewater concentrations of lead and zinc were below chronic water quality criteria (5 and 212 µg/L, respectively) at all sites. However, sculpin, crayfish, and snails were absent from several sites immediately below active mines. Fish densities were also lower at some sites downstream of mines. These results support the findings of previous biomarker and toxicity studies indicating that mining is affecting biota in the Black River drainage near the Viburnum Trend.

INTRODUCTION

Recent proposals to expand areas of prospecting for deposits of lead ore in the Mark Twain National Forest (MTNF) of southeast Missouri raised concerns about potential surface and ground-water quality degradation due to lead mining. In response to concerns about potential impacts in federal lands (MTNF and the Ozark National Science Riverways), the USGS is evaluating the effects of current and proposed mining activities on surface and ground waters in the Missouri Ozarks.

These investigations have found elevated concentrations of lead and other elemental contaminants associated with lead mining. Biochemical effects indicate exposure of fish to environmental lead from sites closest to mines and tailings (Whyte et al. 2002, Allert et al. 2003, and Besser et al. 2003).

STUDY AREA

We studied 12 sites (Figure 1 and Table 1) to characterize the influence of mining activities on metal concentrations and other constituents in surface and pore water, as well as aquatic organisms (fish, benthic invertebrates, crayfish and snails).

Four sites are designated as "reference" sites. These sites were either upstream of mining activity (Site MT3) or in watersheds without mining activity (Sites MT6, MT11, MT12).

Eight sites were designated as "impacted" sites. These sites were located downstream of known mining areas.

OBJECTIVES

- Characterize water quality of surface water and pore water; metal concentrations in surface (Cd, Pb, Zn) and pore water (Cd, Pb, Zn, Ni, Co) and habitat characteristics relative to mining activities.
- Characterize fish and invertebrate communities relative to mining activities and impacts.
- Evaluate relationships between fish and invertebrate communities and water quality, metal concentrations, and habitat variables.

METHODS



Figure 1. Location of sites and mines of the New Lead Belt (Viburnum Trend) in southeast Missouri. Green = Reference; Red = mine impacted; Blue = recovery.

Table 1. Study site locations.

Site No.	Site Description
MT1	West Fork Black River at Satter Bluff
MT2	West Fork Black River at West Fork
MT3	West Fork Black River at Greeley
MT4	Box Fork Creek
MT5	Logan Creek
MT6	Sawewater Creek
MT7	Middle Fork Black River
MT8	Sliding Fork Creek
MT9	South Creek
MT10	Courtois Creek
MT11	Blair Creek
MT12	Hazard Creek



BIOLOGICAL ASSESSMENT

Sampling occurred during September 2003, not yet collected.

Fish samples were collected using two techniques: electrofishing, kick seining (4.5-m², 6-mm mesh seine) in riffle habitats, and sweep seining (6-mm mesh, sweep net) through the entire site.

Grab samples of surface water were collected and analyzed for temperature, pH, conductivity, dissolved oxygen, alkalinity, hardness, ammonia, OH^- , dissolved organic carbon, particulate organic carbon, and sulfate.

Pure water was collected using pipettes and analyzed for temperature, pH, conductivity, dissolved oxygen, alkalinity, hardness, ammonia, dissolved organic carbon, and sulfate.

Micro-habitat measurements included current velocity, depth, particle size of substrate (at location of benthic invertebrate sampling), canopy, embeddedness, and bank stability. Macro-benthic variables included substrate particle size in giles and channel morphology. GIS data included area of watershed and area of mine tailings within watershed.

Three replicate invertebrate samples were collected using a 0.1-m² Hess sampler with a 0.3-mm mesh collection bag in coarse substrate (riffle habitat).

STATISTICAL ANALYSIS

Principal component, correlation and stepwise regression analyses were conducted on the untransformed data (SAS 2000).

METAL ANALYSIS

Metals in filtered (0.45 µm) surface and pore water were analyzed using ICP-mass spectrometry (May et al. 1997). Surface water was analyzed for Cd, Pb, and Zn. Pore water was analyzed for Cd, Pb, Zn, Ni, and Co. Samples were acidified to 1% v/v with ultra-pure nitric acid and analyzed without digestion.

RESULTS AND DISCUSSION

WATER QUALITY

Table 2. Quality of selected surface and porewater variables. Not yet collected.

Site No.	Temp (°C)	pH	Cond (µmhos/cm)	Hard (mg/L CaCO ₃)	Sulfate (mg/L)
Surface water					
MT1	18.8	8.13	366	182	36
MT2	20.1	8.13	414	205	48
MT3	17.8	8.45	344	167	1.6
MT4	17.7	7.85	482	182	79
MT5	18.4	7.88	376	187	58
MT6	17.7	8.05	274	161	24
MT7	21.3	7.20	401	206	88
MT8	21.4	7.42	223	113	8.73
MT9	19.3	7.57	654	452	nd
MT10	18.7	8.11	447	231	48
MT11	17.6	8.14	302	166	nd
MT12	19.8	8.12	455	276	44
Pore water					
MT1	18.5	7.96	378	185	36
MT2	20.8	7.90	412	203	48
MT3	17.3	8.36	340	160	0.57
MT4	17.1	7.98	480	177	79
MT5	18.1	7.71	355	181	46
MT6	18.3	7.78	261	167	17
MT7	22.9	7.68	453	212	91
MT8	21.8	7.48	222	104	8.17
MT9	17.8	7.82	660	460	nd
MT10	18.1	8.03	457	240	52
MT11	17.5	8.04	304	166	nd
MT12	18.3	8.14	445	272	41



Deploying porewater samplers (pipettes) in riffle habitat.

FISH COMMUNITY STRUCTURE

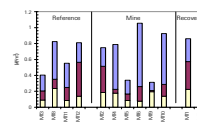


Figure 2. Density of fish for each sampling method, by site. EF = electrofishing; KS = kick seining; SS = sweep seining.

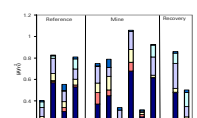


Figure 3. Density (#/m²) of invertebrates collected at 10 sites.

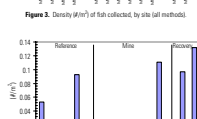


Figure 4. Density (#/m²) of sculpin collected, by site.

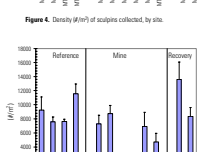


Figure 5. Density (#/m²) of invertebrates collected at 10 sites.

METAL CONCENTRATIONS IN WATER

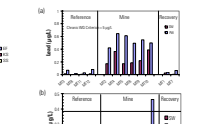


Figure 6. Metal concentrations in surface and pore waters. See text for definitions of site types and abbreviations.

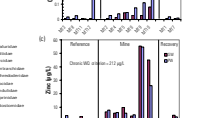


Figure 7. Toxic Units (TU) for surface and porewater metals, based on chronic water quality criteria (WCQ) (USEPA 2002). TU = metal concentration / WCQ.

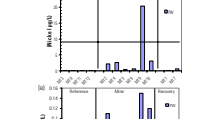


Figure 8. Principal Component Analysis (PCA) plot showing the relationship between fish and invertebrate communities and water quality variables.

Water quality of surface and pore waters from sites downstream of mine tailings were generally elevated in conductivity, hardness, and sulfate (Table 2).

Density of fish collected by each sampling method was similar across sites (Figure 2). Density of fish was generally lower at sites most affected by mining (MT2, MT5, MT6).

The distributions of fish families were generally similar across sites (Figure 3). Cyprinidae was the most common fish family collected by all methods.

The distribution and abundance of sculpins may be impacted by mining (Figure 4). Sites downstream of mining (except for MT10) had few or no sculpins.

Density of invertebrates (Figure 5) was generally lower at mine sites.

Metal concentrations (Figure 6) in pore water were greater than surface water.

Cumulative metal toxic units differed between surface and pore waters (Figure 7). Toxic units exceeded 1.0 only in pore water, and only at sites (MT5, MT10) downstream of mining.

Toxic units were dominated by the influence of cadmium and zinc in surface waters, and by the influence of nickel, cadmium, and zinc in pore waters.

Sites were discriminated using Principal Component Analysis (Figure 8). Mine impacted sites were discriminated primarily by water quality (hard, Ni) in PWA, habitat canopy, and sulfate (>38 mm) and density of Ozark sculpins.

Correlation analysis showed that fish species richness was related to only habitat variables (Table 3), however density of fish and invertebrates were correlated with both habitat variables and water quality. Density of sculpins was only correlated with habitat variables.

Stepwise regression of fish species richness, and densities of sculpins and invertebrates suggested that habitat variables greatly affect fish and invertebrate communities (Table 4).

Table 3. Correlation (r values) of fish, water quality, habitat and GIS variables. EF = electrofishing; KS = sweep seining. TU = toxic units.

Parameter	Fish Species Richness
SS sweep seining	0.25
Area of watershed (mi ²)	0.22
Distribution of particles in giles	0.20

Table 4. Stepwise regression (ascending cumulative R²) of fish, water quality, habitat, and GIS variables.

Parameter	Pearson R	p	R ²
Fish Species Richness	0.25	0.0007	0.06
Area of watershed (mi ²)	0.22	0.0008	0.18
Distribution of particles in giles	0.20	0.0008	0.34
Area of watershed (mi ²)	0.19	0.0008	0.42
Area of watershed (mi ²)	0.18	0.0008	0.58
Area of watershed (mi ²)	0.17	0.0008	0.74
Area of watershed (mi ²)	0.16	0.0008	0.90
Area of watershed (mi ²)	0.15	0.0008	0.96
Area of watershed (mi ²)	0.14	0.0008	0.99
Area of watershed (mi ²)	0.13	0.0008	1.00

Table 5. Stepwise regression (ascending cumulative R²) of fish, water quality, habitat, and GIS variables.

Parameter	Pearson R	p	R ²
Sculpin Density (#/m ²)	0.20	0.0007	0.16
Area of watershed (mi ²)	0.19	0.0008	0.32
Area of watershed (mi ²)	0.18	0.0008	0.45
Area of watershed (mi ²)	0.17	0.0008	0.58
Area of watershed (mi ²)	0.16	0.0008	0.71
Area of watershed (mi ²)	0.15	0.0008	0.84
Area of watershed (mi ²)	0.14	0.0008	0.97
Area of watershed (mi ²)	0.13	0.0008	1.00

Table 6. Stepwise regression (ascending cumulative R²) of fish, water quality, habitat, and GIS variables.

Parameter	Pearson R	p	R ²
Invertebrate Density (#/m ²)	0.20	0.0007	0.16
Area of watershed (mi ²)	0.19	0.0008	0.32
Area of watershed (mi ²)	0.18	0.0008	0.45
Area of watershed (mi ²)	0.17	0.0008	0.58
Area of watershed (mi ²)	0.16	0.0008	0.71
Area of watershed (mi ²)	0.15	0.0008	0.84
Area of watershed (mi ²)	0.14	0.0008	0.97
Area of watershed (mi ²)	0.13	0.0008	1.00

Table 7. Toxic Units (TU) for surface and porewater metals, based on chronic water quality criteria (WCQ) (USEPA 2002). TU = metal concentration / WCQ.

Parameter	Pearson R	p	R ²
Surface TU	0.20	0.0007	0.16
Area of watershed (mi ²)	0.19	0.0008	0.32
Area of watershed (mi ²)	0.18	0.0008	0.45
Area of watershed (mi ²)	0.17	0.0008	0.58
Area of watershed (mi ²)	0.16	0.0008	0.71
Area of watershed (mi ²)	0.15	0.0008	0.84
Area of watershed (mi ²)	0.14	0.0008	0.97
Area of watershed (mi ²)	0.13	0.0008	1.00

Table 8. Stepwise regression (ascending cumulative R²) of fish, water quality, habitat, and GIS variables.



Hazard Creek

CONCLUSIONS

- Metal concentrations, conductivity, and hardness from stream sites near mining activities were increased relative to reference sites. Surface and pore water trends were similar, although there were higher concentrations of metals in pore water.
- Conductivity and hardness may be good surrogate measures for the impacts of mining in Ozark streams.
- Fish communities were affected at sites downstream of mining sites. Sculpins were absent from several sites downstream of mines, which may reflect their sensitivity to metals (Besser et al. 2004). SETAC 2004 Poster Number PH93
- Benthic invertebrate data showed similar trends as fish densities, with reduced density of invertebrates at several sites downstream of mining.
- Physical habitat variables were significantly correlated with fish and invertebrate community indices, therefore, habitat variables, such as depth, velocity, bank stability and embeddedness must be considered when assessing the impact of fish communities from mining.

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